

---

# Deforestation in Brazilian Amazonia: History, Rates, and Consequences

PHILIP M. FEARNSIDE

Instituto Nacional de Pesquisas da Amazônia (INPA), Caixa Postal 478, Manaus 69083-000, Amazonas, Brasil,  
email pmfearn@inpa.gov.br

---

**Abstract:** *Brazil's Amazon forest remained largely intact until the "modern" era of deforestation began with the inauguration of the Transamazon Highway in 1970. Amazonian deforestation rates have trended upward since 1991, with clearing proceeding at a variable but rapid pace. Although Amazonian forests are cut for various reasons, cattle ranching predominates. The large and medium-sized ranches account for about 70% of clearing activity. Profit from beef cattle is only one of the income sources that make deforestation profitable. Forest degradation results from logging, ground fires (facilitated by logging), and the effects of fragmentation and edge formation. Degradation contributes to forest loss. The impacts of deforestation include loss of biodiversity, reduced water cycling (and rainfall), and contributions to global warming. Strategies to slow deforestation include repression through licensing procedures, monitoring, and fines. The severity of penalties for deforestation needs to be sufficient to deter illegal clearing but not so great as to be unenforceable. Policy reform is also needed to address root causes of deforestation, including the role of clearing in establishing land claims.*

Deforestación en la Amazonía Brasileña: Historia, Tasas y Consecuencias

**Resumen:** *El bosque Amazónico de Brasil permaneció prácticamente intacto hasta que comenzó la era "moderna" de deforestación con la inauguración de la Carretera Transamazónica en 1970. Las tasas de deforestación en la Amazonía han tendido a incrementar desde 1991, con desmontes a un paso variable pero rápido. Los bosques Amazónicos son talados por varias razones, pero predomina la ganadería. Los ranchos de mediana y gran extensión fueron responsables de casi 70% de la actividad de desmonte. Las utilidades de la ganadería son solo una de las fuentes de ingreso que hacen rentable a la deforestación. La degradación del bosque resulta de la extracción de maderas, de incendios (facilitados por la extracción de maderas) y los efectos de la fragmentación y de la formación de bordes. La degradación contribuye a la pérdida de bosques. Los impactos de la deforestación incluyen pérdida de biodiversidad, reducción del ciclo de agua (y precipitación) y contribuciones al calentamiento global. Las estrategias para reducir la deforestación incluyen la represión a través de procedimientos de concesión de permisos, monitoreo y multas. La severidad de las penas por deforestación necesita ser suficiente para disuadir la tala ilegal pero no tanto como para ser inaplicable. También se requieren cambios en las políticas para atender de raíz a las causas de la deforestación, incluyendo el papel del desmonte en el establecimiento de posesiones de tierras.*

---

## Introduction

The onslaught on the Amazon began in the early 1970s. Although enormous tracts are still intact, the rate of forest

loss is dramatic, especially in the "arc of deforestation" along the southern and eastern edges. Biodiversity loss and climatic impacts are major concerns, and the vastness of the remaining forests means that the potential

---

*Paper submitted November 16, 2004; revised manuscript accepted January 19, 2005.*

impacts of continued clearing are far more serious than the—already severe—impacts resulting from their loss to date.

Combating deforestation in Brazil is a priority for the government and international organizations. Monitoring and repression is currently the principal strategy. Effective inspection and the levying of fines for those lacking permits from the Brazilian Institute for the Environment (IBAMA), however, must be accompanied by an understanding of the social, economic, and political aspects necessary to address the problem through changes in policy.

### Extent and Rate of Deforestation

By 2003 forest cleared in Brazilian Amazonia had reached  $648.5 \times 10^3 \text{ km}^2$  (16.2% of the  $4 \times 10^6 \text{ km}^2$  originally forested portion of Brazil's  $5 \times 10^6 \text{ km}^2$  Legal Amazon Region), including approximately  $100 \times 10^3 \text{ km}^2$  of "old" (pre-1970) deforestation in Pará and Maranhão (Fig. 1; INPE [Instituto Nacional de Pesquisas Espaciais] 2004). The current rate and cumulative extent of deforestation encompass vast areas. The original extent of Brazil's Amazon forest was approximately the area of Western Eu-

rope. The rate is often discussed in Brazil in terms of "Belgiums"—annual loss approaching the country's area ( $30.5 \times 10^3 \text{ km}^2$ )—whereas the cumulative amount is compared with France ( $547.0 \times 10^3 \text{ km}^2$ ). Almost five centuries of European presence before 1970 deforested an area only slightly larger than Portugal. Current values for deforestation can be obtained from INPE's Web site (<http://www.inpe.br>). The official explanations as to why deforestation rates fluctuate (decrees affecting incentives and programs for inspection and levying fines), however, are unlikely to be correct, as I explain here. In addition, a variety of technical questions about the statistics themselves remain open (Fearnside & Barbosa 2004).

### Causes of Deforestation

In Brazilian Amazonia, the relative weight of small farmers versus large landholders changes continually with economic and demographic pressures. Large landholders are most sensitive to economic changes such as interest rates and other financial returns, government subsidies for agricultural credit, the rate of inflation, and land prices. Tax incentives were a strong driver of deforestation in the 1970s and 1980s (Mahar 1979), and although a decree in 1991 suspended new incentives, the old ones continue, contrary to the popular impression fostered by statements by government officials that all had been ended. Other incentives, such as government-subsidized credit at rates well below inflation, became much scarcer after 1984.

Hyperinflation dominated the economy for decades preceding Brazil's "Plano Real" reform (1994). Land was at a premium, and prices reached levels higher than could be justified as an input to agricultural and ranching production. Deforestation enabled claims to land, and cutting for cattle pasture was the cheapest and most effective in this sense, although the extent to which this activity was land speculation is a matter of debate (Hecht et al. 1988; Faminow 1998; Fearnside 1987, 2002b). Land speculation was important until around 1987, but there was a subsequent increase in the role of pasture profit from beef production (Mattos & Uhl 1994; Margulis 2003).

Brazil's economic recession best explains the decline in deforestation rates from 1987 through 1991. Ranchers were unable to expand their clearings as quickly, and the government lacked funds for road construction and settlement projects. The impact of repressive measures (e.g., helicopter patrols, confiscation of chainsaws, fines) was probably minor. Policy changes on fiscal incentives were also ineffectual. The decree suspending incentives (No. 153) came into effect on 25 June 1991—subsequent to most of the observed decline in deforestation (Fig. 1). Even in that last year (1991) the effect would have been minimal because August was the average date for the Landsat images for the 1991 data set. At the low point in 1991, many ranchers were unable to use their funds for investment in clearing because then-president Fernando

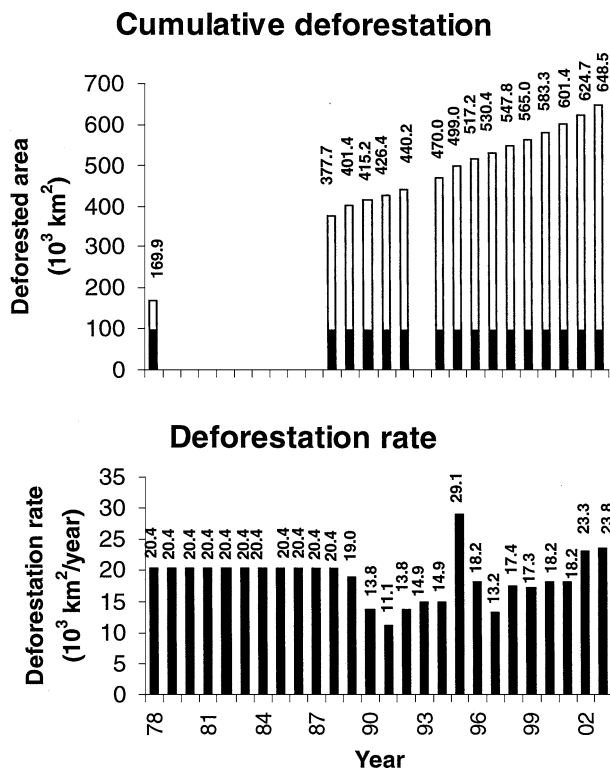


Figure 1. Deforestation in the Brazilian Amazon. For cumulative deforestation the black portion of each bar represents "old" (pre-1970) deforestation. Data from INPE (2004) except for 1978 (Fearnside 1993b).

Collor de Melo had seized bank accounts in March 1990, with funds subsequently released in small installments over a period of years.

The 1995 peak was probably a reflection of economic recovery under the Plano Real. The reforms increased the availability of capital, and municipal elections in 1994 resulted in an increase in agricultural credit. This increase, in providing cash to landholders, has been much more effective in spurring deforestation than economic changes that affected the value of less-liquid assets such as land. The subsequent fall in deforestation rates in 1996 and 1997 was a logical consequence of the Plano Real having sharply cut the rate of inflation. Land values peaked in 1995 and fell by about 50% by the end of 1997. Falling land values made land speculation unattractive. Deforestation rates then climbed to  $17\text{--}18 \times 10^3 \text{ km}^2$  per year, which remained constant for the next 4 years, followed by a jump in 2002 to a new plateau of  $23 \times 10^3 \text{ km}^2$  per year (Fig. 1).

The association of major swings in deforestation rate with macroeconomic factors such as money availability and inflation rate is one indication that much of the clearing is done by those who invest in medium to large cattle ranches, rather than by small farmers who use family labor. The predominant role of larger landholder ranches is evidenced by the location of clearings. Mato Grosso alone accounted for 26% of the  $11.1 \times 10^3 \text{ km}^2$  total in 1991 and had the highest percentage of its privately held land in ranches of  $\geq 1000 \text{ ha}$  (84% at the time of the 1985 agricultural census). By contrast, Rondônia—famous for its deforestation by small farmers—accounted for only 10% of the 1991 total and Acre only 3%. The rise to a rate of  $23 \times 10^3 \text{ km}^2/\text{year}$  in 2002, even with a lackluster domestic economy, can be partly attributed to an increasing globalization of the forces of deforestation, with a sharp growth in the international market for soybeans and especially for beef. Previously, beef had been restricted to the domestic market because of foot-and-mouth disease (Alencar et al. 2004; Kaimowitz et al. 2004).

Understanding who is to blame for deforestation is vital for any program that attempts to reduce it. Surveys conducted in 1998 in the arc of deforestation, which stretches from Paragominas to Rio Branco, found only 25% of the clearing in properties of  $\leq 100 \text{ ha}$  (Nepstad et al. 1999a). The social cost of substantially reducing deforestation rates would therefore be much less than is implied by frequent pronouncements that blame poverty for environmental problems in the region. Thus, strategies such as those that promote agroforestry among small farmers are likely to be ineffectual when cattle ranchers with large estates are the principal villains. Money from drugs, corruption, and many other illegal sources can also be laundered by investing in questionably profitable ventures, such as gold mining dredges and failing cattle ranches. The rapidly increasing drug trade in Amazonia is likely to exacerbate this trend.

The advance of soybean plantations in the region currently poses the greatest threat, with its stimulus for massive government investment in infrastructure such as waterways, railways, and highways. Infrastructure development unleashes an insidious chain of investment and profiteering that can be expected to destroy more forest than the plantations themselves (Fearnside 2001c). Logging roads, especially for mahogany extraction, precede and accompany highways, opening up frontiers for investing timber profits in soybean plantations and cattle ranches. Timber extraction increases the forest's flammability, leading to understory fires that set in motion a vicious cycle of tree mortality, increasing fuel loads, reentry of fire, and eventually total destruction of the forest. What began as undetected deforestation leads to damage that is detectable on Landsat imagery as deforestation (Cochrane et al. 1999; Nepstad et al. 1999b).

Transportation infrastructure accelerates migration to remote areas and increases the clearing of already-established properties. The *Avana Brasil* program, a development package for the period 2000–2007, included US\$20 billion in infrastructure in the Amazon region (Laurance et al. 2001; Nepstad et al. 2001; Fearnside 2002a), mostly driven by the perceived need to transport soybeans. Particularly damaging are the BR-163 (Santarm-Cuiaba) and BR-319 (Manaus–Porto Velho) highways, which can access large blocks of little-disturbed forest. Its successor, the Pluriannual Plan (PPA) for 2004–2007, is virtually identical to *Avana Brasil*.

## The Role of Logging and Fire in Forest Loss

Logging greatly increases the susceptibility of forest to fire. Once fire enters it kills trees and increases fuel loads and understory drying, increasing the risk of more-damaging future fires and the complete degradation of the forest. The impact of the selective logging of low-density, commercially valuable species is often underestimated. The logging process results in the damage of almost twice the volume of the trees being harvested (Verissimo et al. 1992). Because many smaller trees are killed, the effect on individuals is even greater. Near Paragominas, Para, for every tree harvested, 27 trees have been reported killed or severely damaged (Verissimo et al. 1992). Gaps in the canopy allow sun and wind to reach the forest floor, resulting in drier microclimates. The number of rainless days needed for the understory to reach flammable conditions is much less for a forest that has been logged than for one that has not (Nepstad et al. 2004).

In Amazonian forests, fires spread as a slowly moving line of flame in the understory. The bases of many trees are burned as the fire lingers. Amazonian forest trees are not adapted to fire, and mortality from a first burn provides the fuel and dryness needed to make the second

and subsequent fires much more damaging. The temperatures reached and the height of flames in the second fire are significantly greater than in the first, killing many additional trees (Cochrane 2003). After several fires the area is cleared to the point where it appears as deforestation in Landsat imagery (Cochrane et al. 1999; Nepstad et al. 1999b).

During the 1997–1998 El Niño event, the Great Roraima Fire burned 11,394 to 13,928 km<sup>2</sup> of intact primary forest (Barbosa & Fearnside 1999), and fires in the arc of deforestation totaled a further  $15 \times 10^3$  km<sup>2</sup> (Nepstad et al. 1999b; Cochrane 2003). Substantial burning also occurred in logging areas near Tailândia, southern Pará, and in standing forest in the state of Amazonas. In southern Pará, the damage from El Niño events is magnified by a combination of factors: a longer dry season than other parts of Amazonia experience, the concentration of logging activity there, and the concentration of deforestation and associated burning for agriculture and ranching.

## Impacts of Deforestation

### Loss of Productivity

Soil erosion, nutrient depletion, and soil compaction are among the most obvious impacts of deforestation. Agricultural productivity declines as soil quality degrades, although a lower plateau of productivity can be maintained by systems such as shifting cultivation. Continuous inputs of lime, manure, and nutrients can counter degradation, but economic and physical resource limitations render this ineffective for large areas far from urban markets (Fearnside 1997d). Deforestation removes options for sustainable forest management for both timber and genetic and pharmacological resources.

### Changes in Hydrological Regime

Watershed functions are lost when forest is converted to uses such as pasture. Precipitation in deforested areas quickly runs off, creating flash floods followed by periods of greatly reduced or no stream flow. Regular flooding patterns are important for natural ecosystem functioning in and near the river and as for floodplain agriculture.

The percentage of water recycled within the Amazon Basin is now believed to be 20–30% (Lean et al. 1996), rather than the traditional figure of 50% (Salati & Vose 1984). Although indicating that the hydrological impact of deforestation would be less than otherwise thought, in reality the opposite is true. That almost exactly 50% of the rain falling in the basin flows out through the Amazon River implies that the other 50% has been recycled, assuming that water vapor stays within the basin. In fact, some water vapor escapes to the Pacific, espe-

cially in the northwest corner of the basin in Colombia. More importantly, a substantial amount is transported to south and south-central Brazil, Paraguay, Uruguay, and Argentina, and some continues across the Atlantic to southern Africa. This gives Amazonian deforestation a level of impact that is not appreciated at the policy level (Fearnside 2004). Rio de Janeiro and São Paulo were subject to repeated blackouts and electricity rationing in 2001 as a result of low water levels in hydroelectric reservoirs in the non-Amazonian portion of the country.

Water is supplied to central-south Brazil by air currents (low-level jets) coming from Bolivia and from the western part of Brazilian Amazonia (western Rondônia, Acre, and western Amazonas). Water vapor supply to the central-south region has different magnitudes and differing importance depending on the season. During the dry-to-wet transition period (September–October) in southwest Amazonia, water vapor supply is particularly important to avoid a lengthening of the dry season in São Paulo (Brazil's most productive agricultural region). Hydroelectric generation capacity, on the other hand, is particularly dependent on rain in the austral summer (December), corresponding to the rainy season in southwest Amazonia when the difference between the hydrological behavior of forested and deforested areas is least. According to preliminary estimates by Pedro Silva Dias of the University of São Paulo (personal communication), roughly 70% of the rainfall in the state of São Paulo comes from Amazonian water vapor during this period.

In addition to maintaining basin-wide precipitation and long-range water transport, deforestation also produces mesoscale effects. Recent observations of a slight (approximately 5%) increase in rainfall in the heavily deforested Ji-Paraná area of Rondônia, together with satellite observations showing cloud formation occurring preferentially over clearings as small as 5 km in diameter, corroborate preliminary theoretical results on mesoscale effects of deforestation.

The potential of deforestation to increase local precipitation by providing convective updrafts of air that trigger cloud formation might mislead the unwary to conclude that deforestation is not so bad. It could provide a deceptive, temporary improvement as deforestation advances, only to be followed by a precipitous decline in rainfall as deforestation passes a threshold. In addition, the increase of rainfall over a clearing means that rain has been “stolen” from somewhere else. This includes both the distant destinations of water vapor transport and the nearby forest edges. Forest edges would suffer because the convection cells formed over clearings will take wet air aloft to induce rain and will create a downdraft over the nearby forest, bringing dry air down that will inhibit rainfall and dry the forest near the edge of the clearing (perhaps in a band about 20 km wide, provided prevailing winds are not blowing). This drying from edges adds an additional

feedback loop that reinforces the degradation of forest edges through fire and water stress.

### Biodiversity Loss

Biodiversity maintenance is a function to which many attribute value beyond the commercial sale of products (Fearnside 1999). The loss of major portions of Brazil's tropical forests impoverishes the Earth's biodiversity (Capobianco et al. 2001). The biodiversity impact of continued deforestation is much greater in areas with little remaining forest and high levels of endemism, such as the Atlantic Forest. If Amazonian deforestation is allowed to continue to near complete destruction, the same levels of risk to biodiversity will apply there.

### Net Emissions of Greenhouse Gases

Forest fires emit greenhouse gases. The Great Roraima Fire released 17.9 to  $18.3 \times 10^6$  t CO<sub>2</sub>-equivalent C through combustion, of which 67% ( $12.0$  to  $12.3 \times 10^6$  t CO<sub>2</sub>-equivalent C) was from fires in primary forest (Barbosa & Fearnside 1999). Carbon dioxide carbon equivalents are used to compare the emissions from various greenhouse gases based on the global warming potential of each gas over a 100-year time horizon. Clearing at the rate prevailing in 2003 implies approximately  $429 \times 10^6$  tons of CO<sub>2</sub>-equivalent carbon emission, whereas for the 1988–1994 period (the base period used by Brazil for its initial greenhouse gas inventory under the Kyoto Protocol) released  $275 \times 10^6$  tons including all components (updated from Fearnside [2000b], including corrections in Fearnside & Laurance [2004] and Nogueira et al. [2005]), or  $252 \times 10^6$  tons if only the emissions components considered in the National Inventory are used as well as the wood density values available prior to the revision of Nogueira et al. (2005). This figure is slightly more than double the official value of  $116.9 \times 10^6$  tons (MCT 2004: 149). The difference is explained by a series of omitted components in the official estimate (including roots and necromass) and by a high estimate for carbon uptake by secondary forests that does not reflect the slow rate at which these grow in degraded Amazonian pastures.

What most distinguishes the global warming implications of Amazonian deforestation from those of other tropical forests is the huge potential for future emissions. In 1990 net committed emissions from Brazilian deforestation represented 5% of the global total from all sources (including both land-use change and fossil fuels) at that time (Fearnside 1997b), whereas the carbon stock in biomass in Brazilian Amazonia represented 38% of the tropical total (Fearnside 2000a: 129). "Net committed emissions" refers to the net result of emissions and uptakes as an area of forest is replaced by a patchwork of other land uses (in the proportions that would be reached at equilibrium if current land-use patterns continued).

## Strategies to Slow Deforestation

### Repression

In Brazil, deforestation is controlled mainly by repression, through clearing licenses, inspections, and fines. Campaigns are often announced simultaneously with the annual conclusions of INPE's monitoring program. The first major effort to repress deforestation was in 1989 under the Our Nature (*Nossa Natureza*) program. Since then a series of crackdowns has been unsuccessful. Clearing rates in the region seem to rise and fall independent of these programs. Repression, although undoubtedly necessary, needs to be rethought and the underlying causes addressed.

An indication of Brazil's capability to control deforestation came in 2000. After a prohibition on burning became effective in July, imagery from the Advanced Very High Resolution Radiometer (AVHRR) sensor, interpreted at INPE, indicated a drop in fires of more than 80%. Deforestation also decreased because of a deforestation and licensing program that was active in Mato Grosso from 1999–2001, despite subsequent changes in the state government that transformed the program such that it was no longer effective as an impediment to deforestation (Fearnside 2003b; Fearnside & Barbosa 2003).

The reduction of burning in Mato Grosso was achieved by a combination of measures. A system of permits was instituted by the states environment agency (FEMA) that included a printout of a satellite image showing the property boundaries and the existing deforestation. Fines were issued with a satellite image printed on them, thus discouraging argument and attempts to misrepresent the area actually cleared. Portions of Mato Grosso with the greatest decreases in burning were those subjected to special community training and education programs in fire management by the Amazonian Working Group and Friends of the Earth—Brazilian Amazonia, with support from FEMA and from the Prevention and Control Programme for Forest and Savannah Fires in the Legal Amazon. Plans have been announced to extend the system to select municipalities in Pará and Rondônia.

### Policy Reform on Taxes, Credit, and Subsidies

A major problem in controlling deforestation is that much of what needs to be done is outside the purview of agencies in charge of environmental issues. The authority to change tax laws and credit policies rests with other government agencies, as do resettlement policies and road-building and development priorities. Tax subsidies for cattle ranches approved by the Superintendency for Development of Amazonia (SUDAM) were an important force motivating deforestation in the 1970s and 1980s. The cessation of new subsidies in 1991 did not revoke those that

had already been granted. The SUDAM-approved projects gave tax exemption on income they generated and allowed owners to invest in their ranches part of the tax they owed on earnings from operations elsewhere. The exclusion of ranches in 1991 did not affect other damaging activities, such as sawmills and pig-iron smelters fueled by charcoal. The remaining tax subsidies need to be removed.

Another motive for deforestation, more prominent in the 1970s and 1980s than today, is land speculation. The capital gain from selling a property after holding it for a few years was a major source of profit for ranchers when land values increased faster than inflation. Although average land values are no longer increasing at the rates seen before the abrupt slowing of inflation with the 1994 Plano Real, individual properties can still produce speculative profits, particularly when they are near a newly built or improved road. Heavy taxes should be applied to take the profit out of land speculation, both to remove the remaining speculative force in areas favored by infrastructure and to provide protection should there someday be a return to the astronomical inflation rates prevailing in Brazil for most of the last century.

Tax evasion is a significant source of investment funds in Amazonian ranches. Some of the ranchers who deforest the most are medical doctors and other professionals from urban areas. People in such professions often have large incomes that they do not declare. If they were to invest in the stock market or urban real estate they would likely draw the attention of tax authorities, but authorities have little basis for evaluating most of the investment in Amazonian ranches. Even if the soil and rainfall regimes are unfavorable for pasture, resulting in some loss on the investment, money from beef sales from an Amazonian ranch will be "clean." The government must invest in law enforcement and in tightening the tracking of financial movements to eliminate this important driver of deforestation.

Deforestation also receives a strong impetus from subsidized agricultural credit. The government subsidy goes beyond low interest rates and generous grace periods. There are also frequent "amnesties," either forgiving debts or converting them to virtually token payments over long periods at low interest. Amnesties are granted when production is reduced by droughts or other "acts of God." Although usually viewed as one-time interventions, they are a regular feature and represent a large additional subsidy to deforestation.

A variety of other subsidies also increase the profitability of agriculture and ranching. These include price supports for many agricultural products, with government guarantees of the price paid to the farmer, irrespective of how distant the farm may be from markets. Many special programs supply inputs such as fertilizer or lime to specific crops and the vast network of transportation infrastructure, at government expense.

### Land-Tenure and Settlement Policy Reform

The nature of settlements established by the National Institute for Colonization and Agrarian Reform (INCRA) has changed markedly over the years. In the 1970s and 1980s most were placed in areas chosen by INCRA. Since the mid-1990s INCRA has claimed that new settlements are sited only in areas already deforested so as to minimize their impact on deforestation. Despite numerous official statements that such a policy was in effect, new settlements continued to be placed in forested areas, such as the 1996 Rio Acarí and Rio Juma settlements in the state of Amazonas. More recently, INCRA has essentially ceded its role of determining settlement sites to squatter organizations such as the Landless Rural Workers' Movement (MST). Squatters invade either public land or "legal reserves" (areas required to be kept forested) of large ranches, and INCRA subsequently "legalizes" the settlements when they are faits accomplis and compensates the ranchers for the lost land. Because compensation has generally been higher than the market price for land, some ranchers quietly encourage squatters. Bankrupt ranches undergoing foreclosure by the Banco do Brasil have been particularly prone to invasion—a situation that both assures squatters a resistance-free invasion and solves the financial problem of the Banco do Brasil when the compensation is paid by INCRA. The areas chosen by squatters for invasion are invariably under primary tropical forest rather than pasture, agriculture, or secondary forest. The timber provides capital for the squatters, and the soils are considerably better than could be expected in a degraded cattle pasture. The de facto shift of INCRA activity to following in the path of initiatives by landless peasant organizations creates an additional barrier to effective control of this form of deforestation (Fearnside 2001a).

Although small farmers account for only about 30% of deforestation (Fearnside 1993a), its intensity (impact per square kilometer) within the area they occupy is greater than for the medium and large ranchers who hold 89% of the Legal Amazon's private land. Deforestation intensity declines with increasing property size. Deforestation would, therefore, increase if forest areas held by large ranches were redistributed into small holdings. This emphasizes the importance of using already cleared areas for agrarian reform, rather than following the politically easier path of distributing still-forested areas. Large as the area already cleared is, it falls far short of the potential demand for settlement. Indeed, the Legal Amazon as a whole falls short of this demand. Recognizing the existence of carrying-capacity limits, and then maintaining population levels within these limits, is fundamental to any long-term plan for the sustainable occupation of Amazonia (Fearnside 1997c). Deforestation for cattle pasture is considered an "improvement" for the purpose of establishing and maintaining land title. As long as this situation remains, one can expect landholders to clearcut their forest despite

prohibitions. A change in land-titling procedures to cease recognizing pasture as an improvement has yet to take place.

To stem the flow of people to new areas in search of land, agrarian reform is needed in Amazonia and in the areas from which migrants come. The availability of alternative employment in rural and in urban areas is also related to these population flows. At the same time, an “industry of settlement” has grown up, in which people who are granted land in one settlement area sell their claims (often informally if a definitive land title has not yet been granted) and move on to try to obtain a parcel of land in a new settlement. INCRA's often-unsuccessful efforts to detect and disqualify those who have received land before result only in the creation of a permanently landless class that also contributes to deforestation. The goal of providing employment opportunities for all Brazilians will have to be met in ways that are less environmentally destructive than granting plots in Amazonian settlement areas (Fearnside 2001*a*).

### Environmental Services

Economic activities in Amazonia almost exclusively involve material commodities—timber, minerals, agricultural and ranching products, and nontimber products such as natural rubber and Brazil nuts. Finding ways to tap the environmental services of the forest as a means of both sustaining the human population and maintaining the forest (Fearnside 1997*a*, 2003*a*) has much more potential over the long term.

Amazonian forests provide at least three classes of environmental services: biodiversity maintenance, carbon storage, and water cycling. The magnitude and value of these services are poorly quantified, and the diplomatic and other steps through which they might become assets are in their infancy—facts that do not diminish their importance or the pressing need to focus effort on providing the information and the political will required to integrate them into the economy in such a way that maintains rather than destroys the forest. The role of tropical forests in averting global warming is much closer to serving as a basis for international financial flows than are other environmental services such as biodiversity maintenance. This is because the U.N. Convention on Climate Change (UN FCCC) has advanced further than the Convention on Biodiversity, even though both were signed simultaneously at the U.N. Conference on Environment and Development (UNCED) in 1992. The UN FCCC was supplemented with the 1997 Kyoto Protocol.

Investment interest in carbon with a view to short-term returns is likely to be limited, given the fact that the agreement over the Kyoto Protocol reached in July 2001 excludes credit for forest maintenance in the Clean Development Mechanism (CDM) during the protocol's first commitment period (2008–2012). Should this be allowed

in future commitment periods, Brazil could potentially gain substantially from CDM projects to reduce deforestation. A proposal is also under discussion for the creation of a means whereby deforestation avoided now could generate credits after 2012 (M. Santilli et al. 2003. Paper presented at COP [Conference of the Parties]-9. Available from <http://www.ipam.org.br/eventos/cop9/Tropical%20Deforestation%20and%20Kyoto%20Protocol%20COP9.pdf>). The political struggles underlying this Bonn decision on the first commitment period can be expected to shift in the future because the “assigned amount” (national emissions quota) of each party is renegotiated for each successive commitment period, thereby removing the advantage to key actors (especially in Europe) of forcing parties (specifically the United States) to satisfy the commitments they made in Kyoto almost entirely through relatively expensive domestic measures (Fearnside 2001*b*). The negotiations since the 1997 Kyoto conference have been unique because industrialized countries had agreed to specific assigned quotas for the first commitment period before the rules were defined on such questions as inclusion of avoided deforestation in the CDM. For future commitment periods, inclusion of avoided deforestation would help induce countries to agree to larger commitments than they would accept in the absence of such a provision, therefore resulting in a net benefit for climate.

Although not currently favored by Brazil's Ministry of External Affairs, the country always has the option of accepting national limits on emissions that would allow it to earn much more by emissions trading under Article 17 of the protocol, rather than through the CDM of Article 12 (Fearnside 2001*d*). Emissions trading has substantially larger potential for carbon credit because the Kyoto Protocol does not require that the reductions be causally linked to a specific project. It is also not required that the changes be “additional” to what would have occurred in a no-project scenario, the baseline for calculation being the country's first national inventory (i.e., emissions in 1988–1994 in the case of Brazil).

Regardless of the future of decisions on the CDM under the Kyoto Protocol, global warming represents a long-term problem that is likely to gain urgency in the international policy arena as impacts become increasingly apparent to the public and to political leaders. Eventually the major role played by tropical deforestation will be recognized and appropriate measures, Brazilian and international, taken to finance combating deforestation and to provide the basis for an alternative to destructive development.

Many believe the forest will be cut down no matter what and consequently argue that we should worry about other problems. One of the greatest impediments to effective action is fatalism. Fatalism acts as a deterrent to taking action that involves commitment of substantial financial resources and the acceptance of perceived or real

political risks. Many of the key determinants of the future path of development are in the hands of decision makers who need to make their decisions based on the responsibility that this entails. Although the future depends on human decisions, limits also exist. We cannot go on destroying forests without dire and long-lasting consequences.

## Acknowledgments

P.M.F.'s work is supported by the Conselho Nacional de Desenvolvimento Científico e Tecnológico (CNPq: Proc. 470765/01-1) and the Instituto Nacional de Pesquisas da Amazônia (INPA: PPI 1 3620). I thank A. B. Rylands, G. A. B. da Fonseca, K. Brandon, and an anonymous reviewer for comments.

## Literature Cited

- Alencar, A., D. C. Nepstad, D. McGrath, P. Moutinho, P. Pacheco, M. del C. V. Diaz, and B. Soares-Filho. 2004. Desmatamento na Amazônia: indo além da emergência crônica. Instituto de Pesquisa Ambiental da Amazônia, Belém, Brasil (in Portuguese).
- Barbosa, R. I., and P. M. Fearnside. 1999. Incêndios na Amazônia brasileira: estimativa da emissão de gases do efeito estufa pela queima de diferentes ecossistemas de Roraima na passagem do evento "El Niño" (1997/98). *Acta Amazonica* 29:513-534 (in Portuguese).
- Capobianco, J. P. R., A. Veríssimo, A. Moreira, I. dos Santos, L. P. Pinto, and D. Sawyer, editors. 2001. Biodiversidade na Amazônia brasileira (in Portuguese). Editora Estação Liberdade and Instituto Socioambiental, São Paulo. Available from <http://www.isa.org.br/bio/index.htm> (accessed November 2004).
- Cochrane, M. A. 2003. Fire science for rainforests. *Nature*, London 421:913-919.
- Cochrane, M. A., A. Alencar, M. D. Schulze, C. M. Souza Jr., D. C. Nepstad, P. Lefebvre, and E. A. Davidson. 1999. Positive feedbacks in the fire dynamic of closed canopy tropical forests. *Science* 284:1832-1835.
- Faminow, M. D. 1998. Cattle, deforestation and development in the Amazon: an economic and environmental perspective. CAB International, New York.
- Fearnside, P. M. 1987. Causes of deforestation in the Brazilian Amazon. Pages 37-61 in R. F. Dickinson, editor. *The geophysics of Amazonia: vegetation and climate interactions*. John Wiley & Sons, New York.
- Fearnside, P. M. 1993a. Deforestation in Brazilian Amazonia: the effect of population and land tenure. *Ambio* 22:537-545.
- Fearnside, P. M. 1993b. Desmatamento na Amazônia: quem tem razão nos cálculos—o INPE ou a NASA? *Ciência Hoje* 16:6-8 (in Portuguese).
- Fearnside, P. M. 1997a. Environmental services as a strategy for sustainable development in rural Amazonia. *Ecological Economics* 20:53-70.
- Fearnside, P. M. 1997b. Greenhouse gases from deforestation in Brazilian Amazonia: net committed emissions. *Climatic Change* 35:321-360.
- Fearnside, P. M. 1997c. Human carrying capacity estimation in Brazilian Amazonia as a basis for sustainable development. *Environmental Conservation* 24:271-282.
- Fearnside, P. M. 1997d. Limiting factors for development of agriculture and ranching in Brazilian Amazonia. *Revista Brasileira de Biologia* 57:531-549.
- Fearnside, P. M. 1999. Biodiversity as an environmental service in Brazil's Amazonian forests: risks, value and conservation. *Environmental Conservation* 26:305-321.
- Fearnside, P. M. 2000a. Global warming and tropical land-use change: greenhouse gas emissions from biomass burning, decomposition and soils in forest conversion, shifting cultivation and secondary vegetation. *Climatic Change* 46:115-158.
- Fearnside, P. M. 2000b. Greenhouse gas emissions from land-use change in Brazil's Amazon region. Pages 231-249 in R. Lal, J. M. Kimble and B. A. Stewart, editors. *Global climate change and tropical ecosystems. Advances in soil science*. CRC Press, Boca Raton, Florida.
- Fearnside, P. M. 2001a. Land-tenure issues as factors in environmental destruction in Brazilian Amazonia: the case of southern Pará. *World Development* 29:1361-1372.
- Fearnside, P. M. 2001b. Saving tropical forests as a global warming countermeasure: an issue that divides the environmental movement. *Ecological Economics* 39:167-184.
- Fearnside, P. M. 2001c. Soybean cultivation as a threat to the environment in Brazil. *Environmental Conservation* 28:23-38.
- Fearnside, P. M. 2001d. The potential of Brazil's forest sector for mitigating global warming under the Kyoto Protocol. *Mitigation and Adaptation Strategies for Global Change* 6:355-372.
- Fearnside, P. M. 2002a. Avanço Brasil: environmental and social consequences of Brazil's planned infrastructure in Amazonia. *Environmental Management* 30:748-763.
- Fearnside, P. M. 2002b. Can pasture intensification discourage deforestation in the Amazon and Pantanal regions of Brazil? Pages 283-364 in C. H. Wood and R. Porro, editors. *Deforestation and land use in the Amazon*. University Press of Florida, Gainesville.
- Fearnside, P. M. 2003a. Conservation policy in Brazilian Amazonia: understanding the dilemmas. *World Development* 31:757-779.
- Fearnside, P. M. 2003b. Deforestation control in Mato Grosso: a new model for slowing the loss of Brazil's Amazon forest. *Ambio* 32:343-345.
- Fearnside, P. M. 2004. A água de São Paulo e a floresta amazônica. *Ciência Hoje* 34:63-65 (in Portuguese).
- Fearnside, P. M., and R. I. Barbosa. 2003. Avoided deforestation in Amazonia as a global warming mitigation measure: the case of Mato Grosso. *World Resource Review* 15:352-361.
- Fearnside, P. M., and R. I. Barbosa. 2004. Accelerating deforestation in Brazilian Amazonia: towards answering open questions. *Environmental Conservation* 31:7-10.
- Fearnside, P. M., and W. F. Laurance. 2004. Tropical deforestation and greenhouse gas emissions. *Ecological Applications* 14:982-986.
- Hecht, S. B., R. B. Norgaard, and C. Possio. 1988. The economics of cattle ranching in eastern Amazonia. *Interciencia* 13:233-240.
- INPE (Instituto Nacional de Pesquisas Espaciais). 2004. Monitoramento da floresta amazônica brasileira por satélite/Monitoring of the Brazilian Amazon forest by satellite: 2002-2003. INPE, São José dos Campos, São Paulo (in Portuguese).
- Kaimowitz, D., B. Mertens, S. Wunder, and P. Pacheco. 2004. Hamburger connection fuels Amazon destruction. Technical report. Center for International Forest Research, Bogor, Indonesia. Available from [http://www.cifor.cgiar.org/publications/pdf\\_files/media/Amazon.pdf](http://www.cifor.cgiar.org/publications/pdf_files/media/Amazon.pdf) (accessed November 2004).
- Laurance, W. F., M. A. Cochrane, S. Bergen, P. M. Fearnside, P. Delamônica, C. Barber, S. D'Angelo, and T. Fernandes. 2001. The future of the Brazilian Amazon. *Science* 291:438-439.
- Lean, J., C. B. Bunton, C. A. Nobre, and P. R. Rowntree. 1996. The simulated impact of Amazonian deforestation on climate using measured ABRACOS vegetation characteristics. Pages 549-576 in J. H. C. Gash, C. A. Nobre, J. M. Roberts and R. L. Victoria, editors. *Amazonian deforestation and climate*. Wiley, Chichester, United Kingdom.
- Mahar, D. J. 1979. *Frontier development policy in Brazil: a study of Amazonia*. Praeger, New York.
- Margulis, S. 2003. *Causas do desmatamento na Amazônia brasileira* (in Portuguese). The World Bank, Brasília. Available from <http://www.fineprint.com> (accessed November 2004).
- Mattos, M. M., and C. Uhl. 1994. Economic and ecological perspectives

- on ranching in the eastern Amazon. *World Development* **22**:145-158.
- MCT (Ministério da Ciência e Tecnologia). 2004. Brazil's National Communication to the United Nations Framework Convention on Climate Change. General Coordination on Global Climate Change, MCT, Brasília.
- Nepstad, D. C., A. G. Moreira, and A. A. Alencar. 1999a. Flames in the rain forest: origins, impacts and alternatives to Amazonian fires. The World Bank, Brasília.
- Nepstad, D. C., et al. 1999b. Large-scale impoverishment of Amazonian forests by logging and fire. *Nature*, London **398**:505-508.
- Nepstad, D., et al. 2001. Road paving, fire regime feedbacks, and the future of Amazon forests. *Forest Ecology and Management* **154**: 395-407.
- Nepstad, D. C., P. Lefebre, U. L. da Silva, J. Tomasella, P. Schlesinger, L. Solórzano, P. Moutinho, D. Ray, and J. G. Benito. 2004. Amazon drought and its implications for forest flammability and tree growth: a basin-wide analysis. *Global Change Biology* **10**:704-717.
- Nogueira, E. M., B. W. Nelson, and P. M. Fearnside. 2005. Wood density in dense forest in central Amazonia, Brazil. *Forest Ecology and Management* **208**:261-286.
- Salati, E., and P. B. Vose. 1984. Amazon Basin: a system in equilibrium. *Science* **225**:129-138.
- Veríssimo, A., P. Barreto, M. Mattos, R. Tarifa, and C. Uhl. 1992. Logging impacts and prospects for sustainable forest management in an old Amazonian Frontier: the case of Paragominas. *Forest Ecology and Management* **55**:169-199.

